# SOURCES OF VARIATION IN WATERFOWL SURVIVAL RATES 

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#### Abstract

Because of the need to manage hunted populations of waterfowl (Anatidae), biologists have studied many demographic traits of waterfowl by analyzing band recoveries. These analyses have produced the most extensive and best estimates of survival available for any group of birds. Using these data, we examined several factors that might explain variation among annual survival rates to explore large-scale patterns that might be useful in understanding waterfowl population dynamics. We found that geography, body mass, and tribe (i.e. phylogeny) were important in explaining variation in average waterfowl survival rates. Received 22 April 1996, accepted 18 September 1996.


Although modern methods for estimating survival rates have been available for 30 years (Cormack 1964, Jolly 1965, Seber 1965), these estimators have not been applied evenly among all groups of birds. Modern estimation procedures have been put to their fullest use in only one order, the Anseriformes. One reason for this emphasis is the need to understand the relationship between hunter harvest and changes in population parameters (e.g. Anderson and Burnham 1976). Another is that band recoveries reported by hunters provide intensive and extensive samples of birds that otherwise would not be available to researchers. These samples have promoted the development of new estimation procedures and have provided the often "data hungry" methods with sufficient numbers of recoveries for analysis. The estimation of survival rates for many waterfowl species, as well as the thorough documentation of basic biology on those species, provide a unique opportunity for investigating variation in life-history traits. The accessibility of this large data base allowed us to examine factors that might influence average annual survival rates. Previous attempts to understand such relationships (e.g. Zammuto 1986) have been hampered by the use of poor metrics to represent survival rates, e.g. maximum lifespan. Krementz et al. (1989) demonstrated that these metrics should be avoided in such applications.

Our primary interest was to explain variation

[^0]in average annual survival rates based on a variety of ecological, genetic, and evolutionary factors. Previous work in this area has focused on factors such as geography, body mass, phylogeny, behavior, and reproductive effort (Roff 1992, Stearns 1992). We investigated a subset of these factors: geography, body mass, and phylogeny.

Geography.-Analyses of variation in avian survival rates within geographic strata (continents) have demonstrated an effect of location (Krementz et al. 1987, Nichols and Hines 1987, Nichols et al. 1990). One drawback of these studies has been the restricted range (e.g. North America) over which these comparisons have been made. Our data set represents the first sample for birds that covers a wide geographic range (many continents). Laurila (1988) examined variation in age at maturity (which is a correlate of average annual survival rate) in waterfowl across a broad geographic range and concluded that on average, waterfowl with northern breeding ranges began breeding at an older age than waterfowl breeding in temperate and equatorial regions. Gaillard et al. (1989) found that age at first breeding and adult life expectancy tended to covary in birds. Hence, we predicted that in geographic strata more distant from the equator, average annual survival rates should increase, even after controlling for body mass.

Phylogeny.-Phylogenetic history influences many life-history characteristics, including various measures of lifespan (Laurila 1988, Gaillard et al. 1989, but see Brawn et al. 1995). Our sample includes 31 species among 6 tribes in the
order Anseriformes, with an additional 15 cases of replicate estimates for a species/tribe/geographic stratum. After controlling for the effects of body mass on variation in life-history traits, it is possible to examine differences among tribes in average annual survival rates. Previous work along these lines has produced conflicting results. Laurila (1988), Saether (1987, 1989), Gaillard et al. (1989), and Dobson (1990) found that phylogeny covaries with lifespan, whereas Brawn et al. (1995) determined that phylogeny was unrelated to lifespan. This apparent difference may have resulted in part because Brawn et al. (1995) examined only a single order, Passeriformes, whereas the other researchers examined many orders. Based on the findings of Laurila (1988), Saether (1987, 1989), and Gaillard et al. (1989), we predicted that anseriform phylogeny would explain a significant amount of the residual variation in average annual survival rates, after we controlled for body mass. However, we could not predict the direction of these differences.

Body mass. - Body mass relates to many biological characteristics (Western and Ssemakula 1982; Boyce 1984; Saether 1985, 1987, 1989; Laurila 1988; Brawn et al. 1995). Relationships between body mass and phylogeny have suggested that among orders, anseriforms tend to have both late age at first reproduction and high fecundity (Gaillard et al. 1989). Within orders, Saether (1987) found evidence of a "slow-fast continuum" contrasting at the slow end with large eggs, small clutch sizes, long incubation periods, delayed fledging, and delayed maturation, with the opposite traits at the fast end. Thus, we predicted that body mass would explain a large amount of the variation in average annual survival rates, and that waterfowl with larger body masses would have greater survival rates than waterfowl with smaller body masses.

## Data and Methods

We confined our analyses to data from adult females because of their high philopatry to nesting areas (Bellrose 1976) and the probable direct connection between reproductive effort (e.g. egg production) and survival (Roff 1992). When available, we used body mass during the winter to avoid the potentially large fluctuations in mass that occur near the time of egg formation. We used Livezey's $(1986,1991)$ tribal classification.

We used average annual survival rates for adult females except in cases where reported survival rates did not differ between adult females and other age/ sex classes. In several cases, we calculated annual survival rates based on unpublished data and data accessed from the files of the U. S. Fish and Wildlife Service, and the New Zealand Department of Conservation's Banding Office. Unless stated otherwise, the methods of Brownie et al. (1985) were used for recoveries (dead animals) and those of Cormack-JollySeber (Pollock et al. 1990) for recaptures (live animals). When given a choice among models, we chose the most parsimonious model that fit the data adequately. If the chosen model included temporal variation in survival rates, then we used the average survival rate.
In categorizing survival rates by geographic strata, we assigned New Zealand to a separate stratum from Australia because of its disproportionate number of endemic waterfowl species (Marchant and Higgins 1990) and its long geological separation from Australia. To model variation in survival rates among species from various locations, we fitted a linear model to survival-rate estimates. All estimates were obtained using modern banding and capture-recapture models based on Brownie et al (1985) and were maximum likelihood or bias-adjusted maximum likelihood estimates.

Survival-rate estimates were subject to two sources of variation: (1) that arising from the sampling process, and (2) the naturally occurring variation among the true survival rates. Only the second source of variation is of biological interest, but in the modeling process it is important to distinguish between the two sources. Because of this process, standard procedures and software could not be applied. Instead, we used the GAUSS (Aptech 1988) programming language to fit the model described below.

Notation. -The probability that a bird of species $j$, tribe $k$, stratum $l$, survives from the start of year $t$ until the start of year $t+1$ is $S_{i, k, i}$. For convenience, we use the subscript $i(i=1, \ldots, n)$ to denote a unique combination of the subscripts $j, k$, and $l . \hat{S}_{i}$ is an estimate of $S_{i}$, and $\hat{Y}_{i}$ is an estimate of $Y_{i}$, where

$$
\begin{equation*}
Y_{i}=\operatorname{logit}\left(S_{i}\right)=\ln \left[\frac{S_{i}}{1-S_{i}}\right] \tag{1}
\end{equation*}
$$

Model.-The model has two components. The first models the process of obtaining the estimated $Y_{i}$ and the second the relationship among the $Y_{i}$ s. Our interest lies in the second component, i.e. the relationship among the $Y_{i}$ s that contains the information of biological relevance. The first component of the model is required to ensure that the inference process is not corrupted by unmodeled variation arising during the sampling process. The formal model for these two components is as follows.

Conditional on the true value $Y_{i}$, the estimated values $\hat{Y}_{i}$ were modeled as independent normal random
variables with expected value $E\left[\hat{Y}_{i} \mid Y_{i}\right]=Y_{i}$, and variance $\operatorname{Var}\left(\hat{Y}_{i} \mid Y_{i}\right)=\tau_{i}^{2}$. From the estimates $\hat{S}_{i}$ and their standard errors $\operatorname{SE}\left(\hat{S}_{i}\right)$, we obtained estimates of $Y_{i}$ and $\tau_{i}$ using

$$
\begin{equation*}
\hat{Y}_{i}=\ln \left[\frac{\hat{S_{i}}}{1-\hat{S}_{i}}\right] \tag{2}
\end{equation*}
$$

and

$$
\begin{equation*}
\hat{\tau}_{i}^{2}=\left[\frac{\operatorname{se}\left(\hat{S}_{i}\right)}{\hat{S_{i}}\left(1-\hat{S}_{i}\right)}\right]^{2} . \tag{3}
\end{equation*}
$$

From maximum-likelihood theory (Mood et al. 1974), equation 2 is the maximum-likelihood estimate of $\hat{Y}_{i}$ and equation 3 provides a consistent estimate of $\tau_{i}^{2}$.

In the most general model, the $Y_{i} s$ were modeled as independent normal random variables with mean $\mu_{i}$ and variance $\delta_{i}^{2}$. Various hypotheses were tested using restricted versions of this model created by imposing constraints on the $\mu_{i}$ and $\delta_{i}^{2}$.

From equations 2 and 3, and the properties of the normal distribution, it follows unconditionally that the $\hat{Y}_{i}$ values are independent normal random variables with mean $\mu_{i}$ and variance $\delta_{i}^{2}+\tau_{i}^{2}$, and $y=\left(\hat{Y}_{1}\right.$, $\left.\ldots, \hat{Y}_{n}\right)^{\prime}$ is multivariate normal with mean vector $\mu$ $=\left(\mu_{1}, \ldots, \mu_{n}\right)$ and diagonal variance-covariance matrix $\Sigma$ with $i$ th diagonal element $\delta_{i}^{2}+\tau_{i}^{2}$. All of the models we fitted were linear, hence $\mu$ could be expressed in the form $\mu=X \beta$ where $X$ is the model matrix and $\beta$ is a parameter vector.

Parameter estimation.-So far, all of the models considered include $n \tau_{i}^{2}$, and including $n \mu_{i}$ and $n \delta_{i}$, we have more parameters than data. To reduce the number of parameters we make use of the estimates of $\tau_{i}^{2}$ obtained via equation 3 in conjunction with the estimates of $Y_{i}$, substituting these $\hat{\tau}_{i}^{2}$ for the $\tau_{i}^{2}$ in $\Sigma$. Thus our analysis is conditional on these $\hat{\tau}_{i}{ }^{2}$.

Maximum-likelihood parameter estimates were obtained using a two-stage iterative procedure (Draper and Smith 1981). Conditional on $\Sigma$ the maximumlikelihood solutions to $\beta$ are

$$
\begin{equation*}
\hat{\beta}=\left(X^{\prime} \Sigma^{-1} X\right)^{-1}\left(X^{\prime} \Sigma^{-1} y\right) . \tag{4}
\end{equation*}
$$

Given a particular value for $\beta$, the multivariate likelihood function (5)

$$
\begin{align*}
\mathrm{L}(\mathrm{y}, \beta, \Sigma)= & \frac{-\mathrm{p}}{2} \ln (2 \pi)-\frac{1}{2} \ln (|\Sigma|) \\
& -\frac{1}{2}(y-X \beta)^{\prime} \Sigma^{-1}(y-X \beta) \tag{5}
\end{align*}
$$

can then be maximized with respect to $\delta_{i}^{2}$ using standard numerical procedures. Note that $|\Sigma|$ denotes the determinant of the matrix $\Sigma$, and $p$ is the dimension of the vector $\beta$, i.e. the number of parameters comprising the $\beta$ vector. Once a new value for $\delta_{i}^{2}$ has been obtained, and hence a new value for $\Sigma$, the procedure can then be repeated until convergence.
Hypothesis testing. - We considered three types of models.
(1) (G*T* $\left.B M^{*} B M^{2}, \delta^{2}\right)$. In this homoscedastic linear regression model, values of $\mu_{i}$ were modeled as linear functions of effects due to geography ( $G$ ), tribe ( $T$ ), body mass ( $B M$ ), and a quadratic function of $\ln \left(B M^{2}\right)$. We investigated the quadratic function because we were interested in whether there was an asymptotic relationship between survival and body mass. All interaction terms except those including $\ln (B M)$ and its square were included in the model. This model included the homoscedastic constraint $\delta_{1}^{2}=\delta_{2}^{2}=\ldots=\delta_{n}^{2}$. Within this group of models we fitted reduced versions to test the significance of the different effects.
(2) ( $\mu_{i}, \delta^{2}$ ). In this model, each unique combination of species, tribe, and stratum had its own mean. The constraint $\delta_{1}^{2}=\delta_{2}^{2}=\ldots=\delta_{n}^{2}$ was included as in the previous model. This model treated multiple estimates for species/tribe/stratum combinations as replicates. Usually these replicates were obtained from different geographic locations within the particular stratum. In some cases the replicates were obtained at the same location but from a different time period. These location/time effects were assumed to be random and hence incorporated within the $\delta^{2}$ term of the model. The inclusion of this model allowed us to test the validity of the linear regression model. The model was fitted to the subset of data for which replicated observations of species/tribe/stratum combination $i$ existed.
(3) ( $\mu_{i}, \delta_{i}^{2}$ ). In this model, each unique combination of species, tribe, and stratum had its own mean as in the previous model. In addition, each unique combination of species, tribe, and stratum had its own $\delta_{i}^{2}$. Thus, this model relaxed the homoscedastic error assumption. As in the previous model, this model was only fitted to the subset of data for which replicated observations of species/tribe/stratum combination $i$ existed.

Models were compared using the generalized likelihood ratio test (GLRT; Lebreton et al. 1992). To compare ( $G^{*} T^{*} B M^{*} B M^{2}, \delta^{2}$ ) with ( $\mu_{i}, \delta^{2}$ ) and ( $\mu_{i} \delta_{i}^{2}$ ) using the GLRT, the models needed to be nested. For these comparisons, this was ensured by fitting ( $G^{*} T^{*} B M$ * $B M^{2}, \delta^{2}$ ) to the same subset of data used to fit ( $\mu_{i}, \delta^{2}$ ) and ( $\mu_{i}, \delta_{i}^{2}$ ).

## Results

Data from six tribes were collected. Not all tribes were equally represented (e.g. only one species for the Tadornini). The best represented tribe was the Anatini with 14 species from five strata. Annual survival-rate estimates across all species ranged from a low of 0.334 to a high of 0.935 with standard errors of the estimates ranging from 0.010 to 0.294 (Table 1). In all, 107 estimates were collected from five strata. North

Table 1. Average annual survival rate ( $\times 100$ ), standard error (in parentheses), and body mass ( kg ) of adult female waterfowl from around the world.

| Species | Survival | Estimator ${ }^{\text {a }}$ | Reference ${ }^{\text {b }}$ | Body mass | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Africa |  |  |  |  |  |
| Anas undulata | 72.2 (4.48) | B | 1 | 1.008 | Dunning 1993 |
| Australia |  |  |  |  |  |
| Anas gibberifrons | 55.0 (17.0) | B | 2 | 0.474 | Marchant \& Higgins 1990 |
| Anas superciliosa | 63.0 (6.0) | B | 2 | 1.025 | Marchant \& Higgins 1990 |
| Europe |  |  |  |  |  |
| Bucephala clangula | 77.2 (3.4) | C | 3 | 0.787 | Dunning 1993 |
| Somateria mollissima | 89.5 (1.46) | C | 4 | 2.142 | Dunning 1993 |
| Anas clypeata | 58.1 (2.96) | B | 5 | 0.680 | Dunning 1993 |
| Aythya ferina | 59.2 (2.22) | B | 5 | 0.807 | Dunning 1993 |
| Aythya fuligula | 70.7 (1.96) | B | $\begin{gathered} 5 \\ \text { Zealand } \end{gathered}$ | 0.680 | Dunning 1993 |
| New Zealand |  |  |  |  |  |
| Anas platyrhynchos | 44.0 (5.0) | B | 6 | 1.123 | Marchant \& Higgins 1990 |
|  | 41.0 (4.0) |  |  |  |  |
|  | 40.0 (5.0) |  |  |  |  |
|  | 59.0 (6.0) |  |  |  |  |
| Anas superciliosa | 55.0 (12.0) | B | 7 | 0.981 | Marchant \& Higgins 1990 |
| Anas rhynchotis | 68.2 (3.56) | B | 8 | 0.614 | Marchant \& Higgins 1990 |
| Anas gibberifrons | 50.1 (11.29) | B | 8 | 0.469 | Dunning 1993 |
| Cygnus atratus | 84.3 (5.1) | B | 8 | 5.100 | Marchant \& Higgins 1990 |
| Branta c. maxima | 74.0 (1.7) | B | 5 | 4.489 | Marchant \& Higgins 1990 |
|  | 77.0 (1.1) | B | 5 |  |  |
| Tadorna variegata | 79.4 (19.9) | B | 8 | 1.387 | Marchant \& Higgins 1990 |
|  | 55.5 (2.4) |  |  |  |  |
|  | 52.6 (5.7) |  |  |  |  |
|  | 70.1 (3.93) |  |  |  |  |
|  | 64.4 (4.45) |  |  |  |  |
|  | 68.2 (3.56) |  |  |  |  |
| North America |  |  |  |  |  |
| Cygnus buccinator <br> Cygnus columbianus | 80.0 (1.0) | B | 9 | 10.300 | Dunning 1993 |
|  | 92.3 (4.3) | S | 6 | 6.255 | Bellrose 1976 |
|  | 91.9 (4.7) |  |  |  |  |
|  | 93.5 (6.6) |  |  |  |  |
| Anser albifrons | 64.9 (6.9) | B | 10 | 2.456 | Dunning 1993 |
| Chen caerulescens | 81.6 (1.6) | B | 11 | 2.517 | Dunning 1993 |
| Chen canagica | 58.6 (4.5) | C | 12 | 2.150 | C. Dau \& J. Schmutz unpubl. |
| Branta c. canadensis | 66.8 (2.6) | B | 13 | 3.420 | Bellrose 1976 |
|  | 71.9 (1.5) |  |  |  |  |
|  | 68.2 (1.9) |  |  |  |  |
| B. c. occidentalis | 83.6 (4.3) | B | 14 | 3.043 | Bellrose 1976 |
| B. c. leucoparia | 70.0 (2.0) | C | 15 | 1.940 | Dunning 1993 |
| B. c. interior | 78.6 (3.1) | S | 16 | 3.514 | Bellrose 1976 |
| B. c. maxima | 76.4 (5.47) | B | 17 | 4.995 | Bellrose 1976 |
| B. c. moffiti | 63.6 (1.0) | B | 18 | 4.044 | Dunning 1993 |
| B. c. minima | 59.9 (1.5) | C | 19 | 1.264 | Dunning 1993 |
| Branta bernicla | 79.2 (3.64) | B | 20 | 1.230 | Bellrose 1976 |
|  | 83.5 (4.11) | B | 21 | 1.390 | Bellrose 1976 |
| Aix sponsa | 47.6 (1.66) | B | 22 | 0.635 | Dunning 1993 |
|  | 46.9 (1.85) |  |  |  |  |
|  | 52.4 (2.15) |  |  |  |  |
| Anas americana | 58.2 (3.63) | C | 23 | 0.719 | Dunning 1993 |
|  | 58.9 (6.35) |  |  |  |  |
|  | 66.7 (6.01) |  |  |  |  |
|  | 56.0 (6.0) | C | 24 |  |  |
| Anas platyrhynchos | 56.8 (4.3) | B | 25 | 1.098 | Bellrose 1976 |
|  | 58.1 (12.4) |  |  |  |  |
|  | 56.0 (4.1) |  |  |  |  |
|  | 55.0 (2.5) |  |  |  |  |

Table 1. Continued.

| Species | Referen- |  |  |  | Reference |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Anas rubripes | 58.6 (6.9) | B | 26 | 1.100 | Dunning 1993 |
|  | 59.7 (4.8) |  |  |  |  |
|  | 73.5 (13.7) |  |  |  |  |
|  | 51.3 (7.5) |  |  |  |  |
|  | 53.6 (8.8) |  |  |  |  |
|  | 45.4 (3.9) |  |  |  |  |
|  | 48.7 (6.6) |  |  |  |  |
|  | 57.3 (5.2) |  |  |  |  |
|  | 39.1 (5.3) |  |  |  |  |
|  | 36.8 (4.5) |  |  |  |  |
|  | 52.1 (3.8) |  |  |  |  |
|  | 50.6 (2.5) |  |  |  |  |
| Anas acuta | 69.3 (6.0) | B | 27 | 0.860 | Bellrose 1976 |
|  | 62.0 (7.4) |  |  |  |  |
|  | 61.4 (4.5) |  |  |  |  |
|  | 57.3 (5.4) |  |  |  |  |
|  | 60.3 (6.1) |  |  |  |  |
|  | 71.2 (4.0) |  |  |  |  |
|  | 64.7 (2.5) |  |  |  |  |
|  | 56.7 (4.9) |  |  |  |  |
|  | 54.2 (5.1) |  |  |  |  |
|  | 55.1 (4.6) |  |  |  |  |
|  | 76.9 (9.7) |  |  |  |  |
|  | 70.8 (5.4) |  |  |  |  |
|  | 50.0 (4.5) |  |  |  |  |
|  | 42.1 (21.5) |  |  |  |  |
|  | 66.8 (10.6) |  |  |  |  |
|  | 70.1 (5.3) |  |  |  |  |
|  | 57.1 (9.1) |  |  |  |  |
|  | 59.8 (4.1) |  |  |  |  |
| Anas discors | 52.2 (1.8) | B | 28 | 0.374 | Bellrose 1976 |
|  | 50.0 (2.7) |  |  |  |  |
|  | 57.3 (3.1) |  |  |  |  |
|  | 53.1 (9.2) |  |  |  |  |
|  | 41.0 (6.0) | C | 24 |  |  |
| Anas crecca | 68.9 (15.1) | B | 29 | 0.318 | Dunning 1993 |
|  | 50.7 (29.4) |  |  |  |  |
|  | 33.4 (3.22) |  |  |  |  |
|  | 53.1 (6.01) |  |  |  |  |
|  | 46.5 (3.04) |  |  |  |  |
|  | 42.8 (14.98) |  |  |  |  |
|  | 42.9 (4.75) |  |  |  |  |
| Anas strepera | 69.0 (6.51) | B | 30 | 0.828 | Bellrose 1976 |
|  | 56.0 (5.0) | C | 24 |  |  |
| Anas fulvigula | 46.3 (5.6) | B | 31 | 0.952 | P. Gray unpubl. |
| Anas clypeata | 46.0 (5.0) | C | 24 | 0.590 | Dunning 1993 |
| Aythya collaris | 54.7 (7.7) | B | 32 | 0.680 | Dunning 1993 |
|  | 49.5 (4.3) |  |  |  |  |
|  | 47.1 (8.0) |  |  |  |  |
| Aythya valisineria | 56.2 (5.0) | B | 33 | 1.230 | Nichols \& Haramis 1981 |
|  | 56.1 (5.2) |  |  |  |  |
|  | 69.0 (6.9) |  |  |  |  |
| Melanitta fusca | 77.3 (1.76) | B | 34 | 1.467 | G. Dobush unpubl. |
| Somateria mollissima | 87.3 (1.56) | B | 34 | 1.521 | Bellrose 1976 |

[^1]Table 2. Analysis of deviance for the model Geographic Strata * Tribe * Body Mass * (Body Mass ${ }^{2}$ ), $\delta^{2}$.

|  | $\Delta$ deviance ${ }^{\text {a }}$ | df | $P$ |
| :---: | :---: | :---: | :---: |
| Geographic strata | 4.93 | 4 | 0.295 |
| $\ln$ (body mass) | 62.64 | 1 | <0.0001 |
| $\ln$ (body mass ${ }^{2}$ ) | 0.37 | 1 | 0.545 |
| Tribe | 44.87 | 5 | <0.0001 |
| 2-way interactions (linear) | 27.25 | 11 | 0.004 |
| 3 -way interactions (linear) | 8.94 | 1 | 0.003 |
| 3 -way interactions (quadratic) | 0.32 | 1 | 0.571 |

- $\chi^{2}$ approximation.

America had the greatest number of estimates and Africa the fewest. Tribes were not equally represented across strata; the most tribes were represented by North America and the fewest by Australia and Africa (Table 1).

All variables (geographic strata, tribe, body mass) were important in explaining variation in survival rates (Table 2). Although the main effect of strata was not significant ( $\chi^{2}=4.93$, df $=4, P=0.29$ ), there were significant interactions between strata and tribe and strata and body mass. Thus, strata influenced both the intercept and the slope of the survival rate/body mass relationship. In particular, Anserini from New Zealand and both Anatini and Aythyini from Europe had relatively high survival rates for their body masses ( $P<0.05$ ), but the slopes were similar to those of other tribes. In New Zealand, survival rates for Cygnini and Anatini were lower than those of Anserini. Body mass had a noticeably larger effect on survival rates of European Aythyini than on European Anatini or New Zealand Anserini, but the effect of body mass was reduced at higher body masses. That Europe was involved in three interaction terms suggests that the relationship between body mass and survival rate was different than that from the other strata.

Variation in survival rates was affected by tribe ( $\chi^{2}=44.87, \mathrm{df}=5, P<0.001$; Table 2, Fig. 1). Survival rates were highest for Cygnini, followed by Anserini and Mergini (which were equal), then Anatini and Aythyini (which also were equal). Among tribes represented by more than one species, all except for Cygnini exhibited positive relationships between survival and body mass; this effect was greatest for Mergini (Fig. 1). Survival rates of swans were inversely related to body mass, with Cygnus buccinator having relatively low survival for its size. An-


Fig. 1. Relationships between survival rate and body mass in waterfowl tribes. The Tadornini is not included.
derson et al. (1986) obtained several survivalrate estimates for Cygnus buccinator, which varied from 79.6 to $88.2 \%$, and suggested that the true survival rate was "perhaps $90 \%$." We chose to use the lower estimate of $80 \%$ because it was derived from the largest data set. Nevertheless, the relationship between body mass and survival for swans remains open to debate.
In testing for a lack of linear fit, we examined only those cases where we had replicate estimates for a species/tribe/strata combination ( $n$ $=15$ ). In comparing the full regression model with one allowing each of the 15 species/tribe/ strata combinations to have its own mean, we found significant evidence of a lack of linear fit ( $\chi^{2}=31.48, \mathrm{df}=5, P<0.001$ ). This suggests that additional variation in the data exists that was not accounted for by the main effects and the interaction terms that we modeled. This "unaccounted for" variation does not mean that our results led us to incorrect conclusions. Rather, it suggests that we should be cautious in the strength of our inferences. When we further generalized the model to allow each group to have its own variance ( $\delta_{i}^{2}$ ), we found no evidence that the variance changed with the mean $\mu_{i}\left(\chi^{2}=9.86, \mathrm{df}=14, P=0.77\right)$.

We found strong evidence that survival rates depend on body mass $\left(\chi^{2}=62.64, \mathrm{df}=1, P<\right.$
0.001; Table 2, Fig. 2), but the significant quadratic term ( $\chi^{2}=9.15, P<0.001$ ) indicates that the relationship between $\operatorname{logit}\left(S_{i}\right)$ and body mass is nonlinear. Examining the tribe-specific plots of $\log i t\left(S_{i}\right)$ and body mass revealed that for species with sufficient data (Anserini, Anatini, Mergini), survival rate was low when body mass was either low or high, and was high when body mass was intermediate.

## Discussion

Although we examined the effects of geography essentially at the level of the continent, we found evidence that geography was related to waterfowl life histories (as did Laurila [1988]). Laurila (1988) based his analyses on distance from the equator and found that waterfowl breeding in the arctic: (1) were significantly heavier than temperate or tropical species; (2) matured later in life (i.e. live longer); (3) produced young that matured more rapidly; and (4) experienced higher rates of nest predation. We also found that tribes with significant interactions with geography were from strata farther from the equator; Aythyini and Anatini from Europe, and Anserini from New Zealand, had higher survival than would be expected for their body size. We caution that we lacked data on survival rates of waterfowl residing near the equator. Thus, the notion of an apparent increase in survival with increasing distance from the equator may require reassessment if additional data from southern latitudes becomes available (see Brawn et al. 1995). Laurila (1988) also noted that large-sized waterfowl tend to nest in the arctic, have poorly concealed nests, and are prone to high levels of nest predation. He suggested that large body size evolved as a defense against predators and as a means of reducing relative heat loss (sensu Bergmann's rule).

In analyses of the species/tribe/strata combinations, we found that the main effects and interaction terms that we modeled did not account for the substantial variation in the data. In an attempt to address Saether's (1987) hypothesis that ecological traits explain much of the variation in demographic patterns in birds, Pontier et al. (1990) examined the influence of seven ecological traits on various demographic patterns. Their results supported Saether's hypothesis. Furthermore, they determined that the Anseriformes was at one end of the slow-fast


Fig. 2. Relationship between survival rate and body mass in waterfowl. Confidence limits are $95 \%$ intervals around the regression line.
continuum among the nine orders examined. Pontier et al. (1990) found that the anseriforms were wholly classified as exhibiting characteristics of very rapid turnover, i.e. low life expectancy and high reproductive effort. Although we question whether the Cygnini should be characterized as exhibiting very rapid turnover (see Laurila 1988), we concede that ecological traits might help absorb some of the unexplained variation that we noted.

Unlike Brawn et al. (1995), we and others (Saether 1987, Dobson 1990, Gaillard et al. 1989) have found that phylogeny is important in explaining survival rates. Survival rates varied among tribes, with the Cygnini having the highest rates and the Anatini and Aythyini the lowest. That the Mergini had survival rates comparable to the Anserini suggests that phylogeny does not always coincide with orderly changes in survival rates. The Mergini is quite distant phylogenetically from the Anserini (Livezey 1986) and thus should have had lower survival rates than we found.

Body mass was important in explaining lifehistory parameters (see also Western 1979; Peters 1983; Stearns 1992; Saether 1985, 1987, 1989; Gaillard et al. 1989; Pontier et al. 1990; Dobson 1990; Brawn et al. 1995). Our estimate that $44 \%$ of the variation in survival rates was explained by body mass exceeds the $33 \%$ found by Gaillard et al. (1989). Nonetheless, the similarity of these
estimates is surprising and suggests that the factors responsible for the large amount of explained variation in survival rates are few. Too, our finding that for some tribes, survival rates were low at both low and high body masses is intriguing. We predict that as additional data become available, this same pattern will be demonstrated in the tribes for which we had only small amounts of data. But why does this pattern occur? We agree with Peters (1983) and Gaillard et al. (1989) that the biomechanical constraints of flight constitute a strong selective pressure that probably controls the range of body masses in birds. It appears that waterfowl are being penalized with low survival rates at both ends of the range of body masses. Clearly, body mass plays an important role in determining the life-history characteristics of waterfowl.

## Acknowledgments

We thank the many waterfowl banders throughout the world for making this report possible. We are especially grateful to the Nova Scotia Department of Lands and Forests, A. Hutchinson, the Maine Department of Inland Fisheries and Wildlife, the New Brunswick Fish and Wildlife Branch, C. P. Dau, I. A. Schmutz, D. R. Yparraguirre, J. R. Sauer, W. B. Krohn, T. W. Arnold, W. R. Clark, and F. A. Johnson for providing access to banding files and unpublished estimates of survival rates. We thank D. Johnson, T. Nudds, and E. Rexstad for reviewing our manuscript. We also apologize to authors who have published survival-rate estimates for waterfowl not included here; we did our best to include all estimates known to us at the time of our research.

## Literature Cited

anderson, D. R., and K. P. Burnham. 1976. Population ecology of the Mallard. VI. U. S. Fish and Wildlife Service Resource Publication No. 128.
anderson, D. R., R. C. Herron, and B. Reiswig. 1986. Estimates of annual survival rates of Trumpeter Swans banded 1949-82 near Red Rock Lakes National Wildlife Refuge, Montana. Journal of Wildlife Management 50:218-221.
Aptech Systems, Inc. 1988. The GAUSS system, version 2.0. Aptech Systems, Inc. Kent, Washington.
Bellrose, F. C. 1976. Ducks, geese and swans of North America. Stackpole Books, Harrisburg, Pennsylvania.
Boyce, M. S. 1984. Restitution of $r$ - and $K$-selection as a model of density-dependent natural selection. Annual Review Ecology Systematics 15:427447.

Brawn, J. D., J. R. Karr, and J. D. Nichols. 1995. Demography of birds in a Neotropical forest: Effects of allometry, taxonomy, and ecology. Ecology 76:41-51.
Brownie, C., D. R. Anderson, K. P. Burnham, and D. S. ROBSON. 1985. Statistical inference from band recovery data. U. S. Fish and Wildlife Service Resource Publication No. 156.
Caithness, J., M. J. Williams, And J. D. Nichols. 1991. Survival and recovery rates of sympatric Grey Ducks and Mallards in the Auckland Acclimatization District, New Zealand. Journal of Wildlife Management 55:111-118.
Chu, D. S., J. D. Nichols, J. B. Hestbeck, and J. E. Hines. 1995. Banding reference areas and survival rates of Green-winged Teal, 1950-89. Journal of Wildlife Management 59:487-498.
Conroy, M. J., and R. T. Eberhardt. 1983. Variation in survival and recovery rates of Ring-necked Ducks. Journal of Wildlife Management 47:127137.

Conroy, M. J., J. E. Hines, and B. K. Williams. 1989. Procedures for the analysis of band-recovery data and user instructions for Program MULT. U. S. Fish and Wildlife Resource Publication No. 175.
CORMACK, R. M. 1964. Estimates of survival from the sighting of marked animals. Biometrika 51:429438.

Coulson, J. C. 1984. The population dynamics of the Eider Duck Somateria mollissima and evidence of extensive non-breeding by adult ducks. Ibis 126:525-543.
Dean, W. R. J., and D. M. Skead. 1989. Survival and recovery rates of Yellow-billed Ducks. Journal of Wildlife Management 53:119-122.
Dobson, A. P. 1990. Survival rates and their relationship to life history traits in some common British birds. Current Ornithology 7:115-146.
Dow, H., and S. Fredga. 1984. Factors affecting reproductive output of the Goldeneye Duck Bucephala clangula. Journal of Animal Ecology 53: 679-692.
Draper, N. R., and H. Smith. 1981. Applied regression analysis. John Wiley and Sons, New York.
Dunning, J. B., Jr. 1993. CRC handbook of avian body masses. CRC Press, Boca Raton, Florida.
Francis, C. M., M. H. Richards, F. Cooke, and R. F. Rockwell. 1992. Changes in survival rates of Lesser Snow Geese with age and breeding status. Auk 109:731-747.
Gaillard, J.-M., D. Pontier, D. Allaine, J. D. Lebreton, J. Trouvilliez, and J. Clobert. 1989. An analysis of demographic tactics in birds and mammals. Oikos 56:59-76.
Halse, S. A., I. R. James, P. E. Fitzgerald, D. A. Diepeveen, and D. R. Munro. 1993. Survival and hunting mortality of Pacific Black Ducks and Grey Teal. Journal of Wildlife Management 57:42-48.
Hestbeck, J. B. 1993. Survival of Northern Pintails
banded during winter in North America, 195088. Journal of Wildlife Management 57:590-597.

Hestbeck, J. B. 1994. Survival of Canada Geese banded in winter in the Atlantic Flyway. Journal of Wildlife Management 58:748-756.
Johnson, F. A., J. E. Hines, F. Montalbano, III, and J. D. Nichols. 1986. Effects of liberalized harvest regulations on Wood Ducks in the Atlantic Flyway. Wildlife Society Bulletin 14:383-388.
Jolly, G. M. 1965. Explicit estimates from capturerecapture data with both death and immigrationstochastic model. Biometrika 52:225-247.
Krementz, D. G., M. J. Conroy, J. E. Hines, and H. F. Percival. 1987. Sources of variation in survival and recovery rates of American Black Ducks. Journal of Wildlife Management 51:689-700.
Krementz, D. G., J. R. Sauer, and J. D. Nichols. 1989. Model-based estimates of annual survival rate are preferable to observed maximum lifespan statistics for use in comparative life-history studies. Oikos 56:203-208.
Krohn, W. B., and E. G. Bizeau. 1980. The Rocky Mountain population of the Western Canada Goose: Its distribution, habitats, and management. U. S. Fish and Wildlife Service Special Scientific Report-Wildlife, No. 229.
Laurila, T. 1988. Reproductive strategies in waterfowl: The effect of ultimate environmental factors, size and phylogeny. Ornis Fennica 65:4964.

Lebreton, J.-D., K. P. Burnham, J. Clobert, and D. R. Anderson. 1992. Modeling survival and testing biological hypotheses using marked animals: A unified approach with case studies. Ecological Monographs 62:67-118.
Lensink, C. J. 1988. Survival of aluminum and monel bands on Black Brant. North American Bird Bander 13:33-35.
Livezey. B. C. 1986. A phylogenetic analysis of recent anseriform genera using morphological characters. Auk 103:737-754.
Livezey. B. C. 1991. A phylogenetic analysis and classification of recent dabbling ducks (Tribe Anatini) based on comparative morphology. Auk 108:471-508.
Marchant, S., and P. J. Higgins (Eds.). 1990. Handbook of Australian, New Zealand and Antarctic birds, vol. 1. Oxford University Press, Melbourne, Australia.
Mood, A. M., F. A. Graybill, and D. C. Boes. 1974. Introduction to the theory of statistics, 2nd ed. McGraw-Hill, New York.
Nichols, J. D., J. Bart, R. J. Limpert, W. J. L. Sladen, and J. E. Hines. 1992. Annual survival rates of adult and immature eastern population Tundra Swans. Journal of Wildlife Management 56:485494.

Nichols, J. D., and G. M. Haramis. 1980. Inferences regarding survival and recovery rates of winter-
banded Canvasbacks. Journal of Wildlife Management 44:164-173.
Nichols, J. D., and G. M. Haramis. 1981. Sex-specific differences in winter distribution patterns of Canvasbacks. Condor 82:406-416.
Nichols, J. D., and J. E. Hines. 1987. Population ecology of the Mallard. VIII. U.S. Fish and Wildlife Service Resource Publication No. 162.
Nichols, J. D., M. Williams, and T. Caithness. 1990. Survival and band recovery rates of Mallards in New Zealand. Journal of Wildlife Management 54:629-636.
Peters, R. H. 1983. The ecological implications of body size. Cambridge University Press, Cambridge, United Kingdom.
Petersen, M. R. 1992. Reproductive ecology of Emperor Geese: Survival of adult females. Condor 94:398-406.
Pollock, K. H., J. D. Nichols, C. Brownie, and J. E. Hines. 1990. Statistical inference for capturerecapture experiments. Wildlife Monographs No. 107.

Pontier, D., D. Allaine, J. Trouvilliez, J. Clobert, J.-M. Gaillard, and J. D. Lebreton. 1990. Are the demographic patterns of birds related to biological and environmental parameters? Pages 311-320 in Organizational constraints on the dynamics of evolution (G. Vida and J. MaynardSmith, Eds). Manchester University Press, Manchester, United Kingdom.
Ratti, J. T., D. E. Timm, and D. R. Anderson. 1978. Reevaluation of survival estimates for Vancouver Canada Geese: Application of modern methods. Wildlife Society Bulletin 6:146-148.
Raveling, D. G., J. D. Nichols, J. E. Hines, D. S. Zezulak, J. G. Silveira, J. C. Johnson, T. W. Aldrich, and J. A. Weldon. 1992. Survival of Cackling Canada Geese, 1982-1988. Journal of Wildlife Management 56:63-73.
Rienecker, W. C. 1976. Distribution, harvest and survival of American Wigeon banded in California. California Fish and Game 62:141-153.
Roff, D. A. 1992. The evolution of life histories: Theory and analysis. Chapman and Hall, New York.
SAETHER, B.-E. 1985. Variation in reproductive traits in European passerines in relation to nesting site: Allometric scaling to body weight or adaptive variation? Oecologia 68:7-9.
SaEther, B.-E. 1987. The influence of body weight on the covariation between reproductive traits in European birds. Oikos 48:79-88.
SaEther, B.-E. 1989. Survival rates in relation to body weight in European birds. Ornis Scandinavica 20: 13-21.
Samuel, M. D., D. H. Rusch, and S. Craven. 1990. Influence of neck bands on recovery and survival rates of Canada Geese. Journal of Wildlife Management 54:45-54.

SEber, G.A.F. 1965. A note on the multiple recapture census. Biometrika 52:249-259.
Stearns, S. C. 1992. The evolution of life histories. Oxford University Press, New York.
Szymczak, M. R., and E. A. Rexstad. 1991. Harvest distribution and survival of a Gadwall population. Journal of Wildlife Management 55:592-600.
Tacha, T. C., G. F. Martz, and J. Parker. 1980. Harvest and mortality of Giant Canada Geese in southeastern Michigan. Wildlife Society Bulletin. 8:40-45.
Timm, D. E., AND C. P. DaU. 1979. Productivity, mortality, distribution and population status of $\mathrm{Pa}-$ cific Flyway White-fronted Geese. Pages 280-298 in Management and biology of Pacific Flyway geese (R. L. Jarvis and J. C. Bartonek, Eds.). Oregon State University Bookstores, Corvallis.
Trost, R.E. 1987. Mallard survival and harvest rates: A reexamination of relationships. Transactions of the North American Wildlife and Natural Resources Conference 52:264-284.

Western, D. 1979. Size, life history and ecology in mammals. African Journal of Ecology 17:185-204.
Western, D., and J. Ssemakula. 1982. Life history patterns in birds and mammals and their evolutionary interpretation. Oecologia 54:281-290.
White, G. C. 1983. Program SURVIV user's manual, version 1.1. Los Alamos National Laboratory, Los Alamos, New Mexico.
Yparraguirre, D. R. 1982. Annual survival and wintering distribution of Aleutian Canada Geese, 1976-81. M.S. thesis, Humboldt State University, Arcata, California.
Zammuto, R. M. 1986. Life histories of birds: Clutch size, longevity, and body mass among North American game birds. Canadian Journal Zoology 64:2739-2749.

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[^1]:    - B = Brownie et al. (1985); C = Cormack-Jolly-Seber (Pollock et al. 1990); $\mathrm{S}=$ SURVIV (White 1983).
    ${ }^{\mathrm{b}} 1=$ Dean and Skead (1989); $2=$ Halse et al. (1993); $3=$ Dow and Fredga (1984); $4=$ Coulson (1984); $5=$ J. D. Nichols (unpubl. data); $6=$ Nichols et al. (1992); $7=$ Caithness et al. (1991); $8=$ R. J. Barker (unpubl. data); $9=$ Anderson et al. (1986); $10=\operatorname{Timm}$ and Dau (1979); $11=$ Francis et al. (1992); $12=$ Petersen (1992); $13=$ Hestbeck (1994); $14=$ Ratti et al. (1978); $15=$ Yparraguirre (1982); $16=$ Samuel et al. (1990); $17=$ Tacha et al. (1980); $18=$ Krohn and Bizeau (1980); $19=$ Raveling et al. (1992); $20=$ Conroy et al. (1989); $21=$ Lensink (1988); $22=$ Johnson et al. (1986); $23=$ Rienecker (1976); $24=$ T. S. Arnold and W. R. Clark (unpubl. data); $25=$ Trost (1987); $26=$ Krementz et al. (1987); $27=$ Hestbeck (1993); $28=$ USFWS (unpubl. data); $29=$ Chu et al. (1995); $30=$ Szymczak and Rexstad (1991); $31=$ F. A. Johnson (unpubl. data); $32=$ Conroy and Eberhardt (1983); 33 = Nichols and Haramis (1980); 34 = D. G. Krementz (unpubl. data).

